

Fuels of the Southern Appalachian Fire and Fire Surrogate Study following Hurricane Helene
and the Black Cove Wildfire Complex

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ABSTRACT

Wildland fuel loading estimates are used to approximate potential fire behavior when planning and implementing wildland fire management. Managers consider the relationships of fuel volume and structure to wildland fire behavior and resulting ecological responses. As active fire and fuels management becomes more widespread after decades of active fire suppression and limited or infrequent forest management, managers increasingly seek to incorporate prescribed fire for a range of stewardship goals and outcomes. As development in rural areas expands the wildland urban interface, managers may consider alternative fuel treatments in areas where prescribed fire may not be feasible or practical. Initiated in the early 2000s, the National Fire and Fire Surrogate Study (FFSS) compared the effects of prescribed fire to those of fire surrogate treatments on a variety of ecological responses, including fuels and potential fire behavior, at multiple locations across the country. For the southern Appalachian Mountains, the treatments implemented in the study included prescribed fire, mechanical felling (shrubs and midstory), a combination of prescribed fire and mechanical felling, and untreated controls. Over a period of 24 years, sites were repeatedly treated and monitored to determine the efficacy of each treatment in producing desired outcomes. Following 24 years of management and study, the FFSS site at the Green River Game Land, in Henderson and Polk counties, North Carolina, was impacted by two major disturbances over a six-month period – Hurricane Helene in September 2024 followed by a large wildfire in March 2025. From June to July 2025, fuel loading and

structure at the site were measured and assessed to determine how changes in fuels may have differed between the treated areas. Our results suggest that, for the period between 2014 and 2025, decreases in ground fuel (O Horizon) mass differed between the treatments. The decrease was greatest in the control treatment, which did not differ from the mechanical-only treatment, but did differ from the burn-only and mechanical+burn treatment. Fine woody fuel mass decreased in all treatments and did not differ between the treatments. Coarse woody debris only decreased in the burn-only and mechanical+burn treatments, but differences between the treatments were not statistically significant. Currently, the mean total fuel mass range for the control, mechanical-only, and mechanical+burn treatments is 45.32-50.60 Mg ha⁻¹. The burn-only mean total fuel mass is 26.93 Mg ha⁻¹, nearly half of what is present in the other treated or control areas. Understanding the long-term fuel dynamics at this location will provide critical information for the southern Appalachian Mountain region as scientists and managers consider fuel reduction treatments to alter fuel complexes, with and without storm impacts from Hurricane Helene.

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GENERAL AUDIENCE ABSTRACT

In wildland fire ecology and management, the logs, twigs, leaves, and other combustible materials in a landscape are referred to as *fuel*. The amount of fuel, its dryness, and the way that it lays or stands, affects the way that wildfires burn. An area with abundant fuel, or fuel that is densely packed together, is likely to burn more severely than an area with less fuel, or fuel that is more scattered on the forest floor. Managing fuel is crucial to managing fire in a way that makes it less dangerous for people and more beneficial to nature. Fuels treatments are methods of managing fuel. These treatments may include prescribed burning or other methods of altering fuel volume or structure. Prescribed fire is often the preferred method of managing fuels in the eastern US. Relative to other potential treatment options, it is less expensive and may assist in meeting a variety of land management objectives. However, different treatments impact different ecosystems in different ways, therefore treatment effectiveness is often very site-specific. The National Fire and Fire Surrogate Study, initiated in the early 2000s, sought to determine whether alternative treatments could have the same ecological benefits as prescribed fire. The Green River Game Land in western North Carolina was chosen as the sole research site in the southern Appalachian Mountains. It has been repeatedly treated and monitored since 2001. In 2024, the area was impacted by Hurricane Helene, causing significant impacts to parts of the Game Land, including areas within the study site. In March 2025, a large wildfire burned through Green River, impacting the study site to varying degrees. During June-July 2025, fuel loading and

structure were remeasured to determine how the treated areas responded to both Hurricane Helene and the Black Cove Wildfire Complex. The results showed that the amount of fresh and decomposed leaves decreased most in the control treatments, where no management activities had taken place since 2001. However, the decrease in the control was not greater than areas where shrubs and smaller trees were felled, but it was greater than sites that included only prescribed fire or sites that were managed with prescribed fire and felling of the shrubs and small trees. Overall, the amount of fuel currently present in all of the treatments is lowest in the places where prescribed fires were used without the combination of felling. This information is important to consider as individuals respond to Hurricane Helene and plan for what may be done to reduce wildfire hazard long-term in the southern Appalachian Mountains as managers proceed with future forest restoration efforts or seek to decrease fire hazard in the wildland-urban interface.

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1 **Chapter 1. Introduction**

2 Wildfire ecology of the eastern US remains relatively understudied as compared to the
3 western US, where larger and more intense wildfires draw more national attention (Hiers et al.
4 2020). While wildfires in the East are generally smaller, less intense, and typically less
5 threatening to humans and other natural resources, they can occur at equal or greater frequency
6 than those in the West and may be no less impactful ecologically (Donovan et al. 2023). Fire
7 return intervals are generally shorter in the southeastern US than in most other regions of the
8 country (Gabbert 2015). In the southern Appalachian Mountains, humans’ relationship with
9 wildland fire is long and storied. It pre-dated European settlement and profoundly shaped the
10 region’s ecology (Lafon et al. 2017).

11 North Carolina’s Green River Game Land (GRGL) hosts southern Appalachia’s sole Fire
12 and Fire Surrogate Study (FFSS) research site. One of only four FFSS sites that have remained
13 active since the study’s inception in 2001, and the only active site east of the Mississippi River,
14 GRGL is a crucial source of information for scientists and other professionals seeking to inform
15 research or strategy in wildland fire and fuels management.

16 Since 2002, GRGL has been actively managed in accordance with the FFSS suite of
17 treatments. Mechanical (chainsaw) felling of the shrubs and midstory has occurred twice in the
18 mechanical-only treatment (2002 and 2012) and once in the mechanical+burn treatments (2002).
19 Prescribed burning has occurred five times (2003, 2006, 2012, 2015, and Spring 2024) in the
20 burn-only and mechanical+burn treatments. Numerous research papers have been published to
21 describe the effects of these treatments on ecological properties and processes, with Bernal et al.
22 (2025) and Taylor et al. (2025) serving as the latest assessments prior to the inception of this
23 thesis project.

24 In late September 2024, GRGL was impacted by Hurricane Helene. Its high winds and
25 excessive rain caused extensive impacts to the region (Dale et al. 2025). Within GRGL, the
26 storm caused significant but patchy blowdowns, adding substantial fuel loads in some areas,
27 whereby having little or no discernable impact to others. In impacted areas, downbursts felled
28 patches of mature overstory trees, creating canopy gaps of various sizes and depositing fuels of
29 various timelag classes. Moreover, Helene's arrival coincided with the start of North Carolina's
30 fall wildfire season, adding complexity to incident response by increasing fuel loading and
31 limiting access to vital infrastructure.

32 In late March 2025, the Black Cove Wildfire Complex burned through GRGL, impacting
33 approximately 60% of the FFSS treatments. Storm-related increases in fuel load, in addition to
34 challenges to access posed by damage and blockage of key infrastructure, may have strongly
35 influenced fire effects. Wildfire managers working the fire noted elevated fire behavior in areas
36 that contained what was deemed to be storm-related fuel loading, and many suppression
37 resources were delayed in their response as fire crews spent hours clearing roads and trails of
38 storm debris to gain access for suppression activities.

39 These compounded disturbances presented a unique opportunity to examine the efficacy
40 of the FFSS fuel reduction treatments following two major natural disturbances. In this thesis,
41 Chapter 2 presents a summary of scientific literature related to disturbance ecology and the
42 responses of wildland fuel loads to disturbances in the eastern United States. Chapter 3 presents
43 research from fuels data that was collected at GRGL in Summer 2025 following both Hurricane
44 Helene and the Black Cove Wildfire Complex near Hendersonville, NC. In Chapter 4, I provide
45 next steps and some additional preliminary analyses that should be considered as land managers

46 and scientists move forward in post-Helene forests, particularly if lands they are managing
47 include multiple prescribed fires or fuel reduction treatments.

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64 Chapter 2. Literature Review

65 2.1 History of fire in the southern Appalachian Mountains

66 Fire history records constructed from fire scar and soil charcoal data indicate that
67 southern Appalachian forests experienced frequent low- and mixed-intensity fires for millennia
68 (Lafon et. al 2017). This fire frequency has been drastically altered by human activity in recent
69 years (Alexander et al. 2021, Nowacki and Abrams 2008). In the summer of 1910, the newly-
70 created US Forest Service reeling from the impacts of the largest and most devastating wildfires
71 in the nation’s history began to formulate its approach to fire protection (Pyne 1981). The long
72 legacy of what would come to be known as prescribed fire, used for centuries by indigenous
73 peoples and frontiersman alike (Stambaugh et al. 2018), was strongly curtailed by the US Forest
74 Service in an effort to appease industrial foresters and a general public that had come to fear any
75 and all wildland fire. Following the logging boom of the early 20th century, higher incidence of
76 high intensity fires, many burning through areas heavily loaded with logging slash, prompted
77 authorities to adopt more heavy-handed policies of fire suppression without complimentary
78 programs to manage the accumulation of fuels (Busenburg 2004). In 1935, the US Forest Service
79 established its “10 AM Policy,” mandating that all fires be extinguished by 10 AM the following
80 day. The policy embodied federal and state governments’ inclination to attempt to entirely
81 exclude fire from the landscape. While this policy and others like it failed to completely exclude
82 fire, it succeeded in drastically reducing the frequency and spatial extent of large fires (Lafon et
83 al. 2017).

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85

86 2.2 Modern fire ecology and fuels management

87 In the 1970s, fire management policy in the United States began to shift toward a more
88 sustainable, science-based approach (Pyne 1981). The US Forest Service began to recognize the
89 role of fire in regulating ecosystem function and promoting biodiversity and gradually adopted a
90 “let-burn” policy, allowing naturally ignited fires to burn in designated wilderness areas (Center
91 for Public Lands 2023). After decades of heavy public relation campaigns led by the federal
92 government in support of fire exclusion and suppression, let-burn was not always well received
93 by a public exposed to decades of fire suppression directives and approaches. Yet, while the
94 public’s perception of fire itself may have remained generally negative, that perception likely
95 reflected a healthy fear of dangerous, intense wildfires, rather than the effects produced by the
96 low-intensity “light burning” encouraged by ecologists. Taylor and Daniel (1984) showed
97 members of the general public images depicting various fire effects ranging from low to high
98 severity. Their results showed that participants favored the effects of low intensity prescribed fire
99 over the severe burns caused by wildfires. The results of the study may better reflect public
100 perception of fire in the western US, where the surveys were conducted, than in the southeast,
101 where cultural burning has persisted to a greater degree despite government restrictions.

102 Over the ensuing decades, “good fire” has gradually become better understood and
103 accepted by professional fire managers and the public alike (Wu et al. 2022). Today, prescribed
104 fire is widely viewed as good, healthy, and necessary in many regions and forest types, and is
105 being implemented more frequently. As the role that fire plays in all natural processes is better
106 understood, and as professional and private use of this powerful tool becomes more prevalent,
107 practitioners increasingly seek to refine and adapt fire use to the unique conditions,
108 circumstances, and objectives of their respective management areas.

109 2.3 Fuel loading and fire behavior

110 The accumulation of fuel continues to pose significant challenges for prescribed fire
111 practitioners, firefighters, and ecologists alike (Keane 2015). Fire behavior prediction is
112 dynamic, combining numerous variables and indices with equal parts perception and intuition
113 (Monadero et al. 2019). Fuel characteristics, and the indices derived therefrom, play critical roles
114 in accurately predicting fire behavior (Linn et al. 2020). Understanding and interpreting fuel
115 models is key to evaluating the fuel characteristics of one's own area of responsibility. No
116 existing fuel models account for storm-related fuel loading in the southern Appalachian
117 Mountains (Coates et al. 2019). When faced with large scale post-disturbance fuel loads, as seen
118 in western North Carolina following Hurricane Helene, managers are left to estimate fuel loading
119 based on estimates calculated under normal conditions in their area. Estimates that can be
120 produced using the data collected for this thesis project may be valuable to those seeking to
121 mitigate fuel-related fire hazards in storm-impacted areas.

122 2.4 Deadwood

123 Downed woody material of various sizes, referred to here as deadwood, often constitutes
124 the majority of available fuel mass in a forest (Forrester et al. 2023). The concentration,
125 condition, and structure of these fuels have a significant influence on fire behavior and effects
126 (Keane 2015). The interplay of deadwood characteristics is complex, and thorough assessment
127 and understanding of deadwood dynamics in one's area of responsibility is crucial to effective
128 fire and ecological management. In the relatively frequent, low intensity fire regimes of the
129 southern Appalachian Mountains, fires are less likely to fully consume large-diameter deadwood
130 as compared to those in drier Western ecosystems which experience less frequent but higher
131 intensity wildfires (Evans 2012). In addition, mixed severity fires in the region may ultimately

132 increase deadwood volume where fire-induced mortality adds more material than is consumed
133 (Bernal et al. 2025).

134 2.5 The Fire and Fire Surrogate Study

135 While prescribed fire may often be the most effective, inexpensive, and least destructive
136 tool available to land managers, it may not be feasible in all applications or environmental
137 settings (Waldrop and Goodrick 2012). Acceptance of prescribed fire may vary by region, with
138 regional acceptance or apprehension stemming largely from its perceived threat to various
139 interests. In much of the eastern US, the public may be less amenable to fire use within, or in
140 close proximity to, the wildland-urban interface (WUI). In 2000, the National Fire and Fire
141 Surrogate Study (FFSS) sought to evaluate the impacts of a standardized suite of prescribed fire
142 and fire surrogate treatments on a variety of ecological responses in various ecosystems across
143 the United States (Witherspoon et. al. 2012). The results of this long-term study would, among
144 other things, serve to inform management strategy for areas where fire may be less practical as a
145 tool. In FFSS, a total of 13 research sites were developed at the inception of the study, with
146 continued management and evaluation occurring at four of the sites to date (Bernal et al. 2025).
147 At each site, multiple replicates of four treatments were installed: prescribed fire, site-specific
148 mechanical treatment, prescribed fire combined with mechanical treatment, and untreated
149 controls. Multiple response variables were then measured at each site, including but not limited
150 to vegetation response, macroinvertebrate activity, soil chemistry, fuel loading, and predicted fire
151 behavior. For sites in the western US, fire behavior was predicted using the Fire and Fuels
152 Extension of the Forest Vegetation Simulator (FFE-FVS) (Bernal et al. 2025). This western-
153 centric fire behavior model has not yet been utilized for the Green River Game Land (GRGL) in
154 North Carolina to understand its utility or effectiveness therein. Following the impacts of

155 Hurricane Helene and a large wildfire that occurred two decades after the initial installation of the
156 treatments, GRGL presents a unique opportunity to assess fire behavior and burn severity.

157 2.6 Major storms in the southern Appalachians

158 While tropical and extratropical storms are not unheard of in the southern Appalachian
159 Mountains, they are significantly less frequent and/or impactful there than in the Coastal Plain.
160 Since 1850, Polk and Henderson Counties, North Carolina, have been impacted by just 39 of
161 these storms, while more than 150 storms have impacted coastal Craven County, NC (NOAA
162 2026, <https://coast.noaa.gov/hurricanes>. Accessed: 25 March 2026). Though infrequent and
163 generally understudied, major storms in the region can have significant impacts on forest
164 structure, composition, and ecology from high wind and precipitation (McNab et al. 2004). In the
165 past 40 years, eight major storms in the region – Hugo in September 1989 (Doggett 1993), Opal
166 in October 1995 (Greenberg and McNab 1998, Clinton and Baker 2000), Frances, Jeanne, and
167 Ivan in 2004, and now Helene – have resulted in significant windthrow, altering forest structure
168 and composition (Greenberg 2021). The effects of the interaction between wind disturbance and
169 fire are less understood, but available research suggests that, in addition to potential increases in
170 fuel load, the aforementioned changes in structure and composition can have significant
171 influence on fuel dynamics and potential fire behavior (Cannon et al. 2014).

172 2.7 Managing post-storm fuel loads

173 Hurricanes can drastically alter both the structure and composition of forests (Cannon et
174 al. 2023). Reduction in overstory density of mature forests increases light penetration to the
175 forest floor, creating conditions that favor shade-intolerant species (Diaz et al. 2021). Methods of
176 forest management, including prescribed fire, may be hampered by the volume of heavy fuels
177 present following a major storm, further complicating vegetation dynamics and compounding

178 fuel loading issues for managers seeking to mitigate hazardous fuel accumulation. Pre-emptive
179 forest management for major natural disturbances may forestall these negative impacts.

180 2.8 Summary

181 Evidence gathered by a variety of methods suggests that, from pre-settlement times
182 through the mid-20th century, the southern Appalachian Mountains experienced regular, low
183 intensity fires (Lafon et al. 2017). An uptick in higher severity fires following the logging boom
184 of the late-19th and early-20th century, which threatened the timber supply conservationists
185 sought to protect, prompted a widespread suppression policy which sought to exclude wildfire
186 entirely. The result of this policy was widescale alteration of forest structure and composition,
187 and a decades-long buildup of wildland fuels.

188 Wildland fuel loading is a key factor in potential wildfire intensity and burn severity
189 (Keane 2015). While fuel structure and loading have been studied extensively, as with most
190 things wildfire-related, the bulk of their study has been focused on fuel types of the western
191 United States, where most high intensity wildfires occur (Hiers et al. 2020). Further study of
192 fuels dynamics in the eastern United States, and the Southeast in particular, is needed.

193 The knowledge gleaned from this thesis project should be useful to Appalachian forest
194 and fire managers. Silviculturists and ecologists benefit from a more thorough understanding of
195 the impacts of fuels on fire effects and numerous ecological responses, and fire suppression
196 managers require accurate fuel volume and mass estimates to predict fire behavior. In 2001, the
197 National Fire and Fire Surrogate Study sought to evaluate the efficacy of fire and fire surrogate
198 treatments in achieving multiple ecological objectives, including fuels and fire behavior, in
199 various ecosystems across the United States (Bernal et al. 2025). In some cases, prescribed fire
200 alone may be impractical or ineffective in mitigating hazardous fuels or achieving specific

201 management goals where forests have been drastically altered by the historic suppression of fire.
202 While generally uniform in nature, treatments were sometimes adapted to local conditions. At
203 GRGL, mechanical felling operations were tailored to local cover types, removing ericaceous
204 shrubs that benefitted from historic fire suppression (Waldrop et al. 2016). In addition to its
205 historically regular fire return interval, the southeastern US also experiences regular tropical
206 cyclones. In the 180 years of recorded weather data, the region surrounding Green River has
207 been impacted by 39 such storms (NOAA 2026, <https://coast.noaa.gov/hurricanes>. Accessed: 25
208 March 2026). These impacts now exacerbate the suppression policy-related fuel buildup, posing
209 additional challenges for managers. While these storms are relatively common, as compared to
210 other regions of the US which experience regular wildfires, the interaction between storms and
211 fuels is not well understood.

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234 **Chapter 3. Post-Hurricane Helene Wildfire Fuels of the Fire and Fire Surrogate Study at**
235 **Green River Game Land, North Carolina, USA**

236

237 3.1 Introduction

238 Following a period of frequent, extensive, and intense wildfires during the logging boom
239 of the early-20th century, a policy of fire exclusion was adopted in the United States, in part to
240 protect and support forest regeneration but to also reduce property losses and deaths resulting
241 from wildfires (Pyne 1981, Schullery 1989, Lafon et al. 2017). Prescribed and cultural burning
242 were often restricted in many places, and government organizations sought to suppress wildfires
243 as quickly as possible through direct attack (Pyne 1981). In the eastern US, these suppression
244 practices resulted in widescale alteration of forest composition, structure, and fuels that continue
245 to impact forest management today (Nowacki and Abrams 2008, Vaughan et al. 2022). Fire
246 exclusion along with reduced light resources and high white-tailed deer (*Odocoileus virginianus*)
247 populations, contributes to a process known as “mesophication” (Nowaki and Abrams 2008,
248 Kreye et al. 2013). In the southern Appalachian Mountains as fire-adapted species, such as oaks
249 (*Quercus* spp.) and pines (*Pinus* spp.), face replacement by mesophytic competitors that are no
250 longer constrained by frequent low to medium intensity fires (Waldrop et al. 2016).

251 After decades of mesophication, southern Appalachian Mountain forests are dense and
252 contain large quantities of wildland fuels that are often continuous, both horizontally and
253 vertically (Vaughan et al. 2021). Forests whose stand structure previously had been patchy and
254 heterogeneous prior to industrial logging’s spread into the Southeast in the 1920s have become
255 vastly more even-aged and homogenized (Lydersen 2013). With increased tree density and fuel
256 continuity, both important drivers of fire behavior (Keane 2015), comes the increased risk of
257 high intensity and severity wildfires at longer intervals in lieu of low severity fires at shorter
258 intervals. This risk has been actualized in several fire seasons within the last decade, particularly

259 during the southeastern US's Fall 2016 fire season, when the region experienced 51% of the
260 nation's wildfire ignitions, a total of over 34,000 for that year, burning approximately 647,497 ha
261 across 13 states (NICC 2016).

262 In response to the increased risk of high intensity and severity wildfires throughout the
263 country, the National Fire and Fire Surrogate Study was launched in 2000 to determine the
264 effects of prescribed fire and other potential fuel reduction practices, such as commercial
265 thinning and mechanical felling, on a host of ecosystem properties and processes (Schwilk et al.
266 2009). Thirteen sites were selected for this study, including one in the southern Appalachian
267 Mountains near Hendersonville, North Carolina. Situated along the Green River on the North
268 Carolina Game Commission's Green River Game Land (GRGL), this site presented not only
269 potentially hazardous wildland fuel conditions, but it also faced many of the forest management
270 issues of declining oak and pine regeneration and mesophication favoring less commercially and
271 wildlife-valuable tree species that are consistent throughout the region. Treatments at this site
272 were first conducted in 2001, and subsequent treatments were conducted through May 2024.
273 These treatments included: mechanical (chainsaw) felling of shrubs and midstory stems less than
274 10 cm diameter at breast height (DBH) (M), prescribed burning (B), a combination of
275 mechanical felling and prescribed burning (MB), and an untreated control (C) (Stephens et al.
276 2009).

277 Many refereed journal articles have been produced to highlight the impacts of these
278 treatments over time at GRGL, including Bernal et al. (2025) and Taylor et al. (2025). Bernal et
279 al. (2025) evaluated fuel load changes in ground fuels (GF), fine woody debris (FWD) (0-7.5 cm
280 diameter), and coarse woody debris (CWD) (≥ 7.6 cm diameter) from 2001-2014 suggesting that
281 the greatest alteration to fuel loading was an increase in CWD in the MB treatment. This fuel

282 class increased during the 13-year period, largely due to delayed overstory mortality resulting
283 from the use of multiple low to moderate severity prescribed fires. Taylor et al. (2025) evaluated
284 tree responses to these treatments from the onset of the study through 2018. The authors found
285 that desired tree species regeneration was most positively improved with the MB treatment.

286 In September 2024, Hurricane Helene produced unprecedented impacts throughout parts
287 of the southern Appalachian Mountains (Dale et al. 2025). In addition to widespread historic
288 flooding, in North Carolina alone, hurricane and near-hurricane force winds far inland toppled
289 approximately 332,652 ha of timber, generating an exponential increase in both wildland fuel
290 loading and vertical continuity across the region (FEMA 2025, Dale et al. 2025). Windthrow and
291 stem breakage placed a large volume of mature stems on the ground just prior to the start of the
292 Southeast's Fall wildfire season. These fuels, along with other wind-, rain-, and flood-related
293 damage to infrastructure, limited access to remote areas throughout the region, setting the stage
294 for potentially catastrophic wildfires. Subsequently, these conditions and a prolonged dry period
295 following Hurricane Helene resulted in numerous wildfires throughout the region in March
296 2025. The southern Appalachian Fire and Fire Surrogate Study site was one of the locations
297 impacted by both Hurricane Helene and the March 2025 wildfires, most notably the Black Cove
298 Wildfire Complex (NCAGR, 2025. [https://www.ncagr.gov/news/press-](https://www.ncagr.gov/news/press-releases/2025/04/03/daily-update-black-cove-complex)
299 [releases/2025/04/03/daily-update-black-cove-complex](https://www.ncagr.gov/news/press-releases/2025/04/03/daily-update-black-cove-complex). Accessed 25 March 2026) (Figure 3.1).

300 Due to the intensity of both management and research at GRGL and the unprecedented
301 combination of disturbances at the site from Fall 2024 - Spring 2025, we conducted research in
302 Summer 2025 to better understand how both Hurricane Helene and the Black Cove Wildfire
303 Complex impacted the fuels of the southern Appalachian Fire and Fire Surrogate Study. We
304 formed several hypotheses: 1. Overall, Hurricane Helene and the Black Cove Wildfire Complex

305 increased total fuel mass in all treatments from 2014-2025; 2. Change in fuel loading from 2014-
306 2025 was greatest in the control (C) treatment due to the presence of greater fuel loads in the
307 absence of active treatments since 2001; 3. The MB treatment displayed the least change in fuel
308 loading from 2014-2025; 4. Wildfire spread was reduced in the burn-only (B) and MB treatments
309 due to the prior reduction in GF. This research provides critical information for other portions of
310 the Appalachian Mountains that have similar management histories and similar impacts from
311 both Hurricane Helene and recent wildfires.

312 3.2 Methods

313 3.2.1 Study location

314 The North Carolina Wildlife Resources Commission (NCWRC) manages approximately
315 800,000 ha for hunting, fishing, and recreation under the state's Game Lands Program
316 (NCwildlife.gov, 2026). This includes GRGL, which covers 6,159 ha in NCWRC's Mountain
317 Region across Henderson and Polk counties, North Carolina, USA (Figure 3.2). Green River was
318 selected as the location of the southern Appalachian Mountain research site because it was
319 representative of a forest with a historically short-interval, low to moderate severity fire regime
320 and a high risk of uncharacteristically severe wildfire (Witherspoon et al. 2000).

321 At GRGL, elevation range is 366-793 m and the topography is mountainous. The climate
322 is warm continental with mean annual precipitation of 1,638 mm. Mean annual temperature is
323 17.6°C. Soils at the study site are moderately deep, well-drained Ultisols (Typic Hapludults),
324 primarily from the Evard (Replications 1 and 2) and Clifffield (Replication 3) soil series (Dukes
325 et al. 2020). At the time of study implementation in 2001, forest stand ages were approximately
326 80-120 years (Waldrop et al. 2010, Keenan 1998). Fire history of the region suggests a mean fire
327 return interval of 10 years prior to 1940 (Harmon 1982), however no evidence was present of
328 previous wildfires or agricultural activity at GRGL (Waldrop et al. 2010).

329 Forests of GRGL are mixed oak-pine. On dry, xeric ridges, pitch pine (*Pinus rigida* Mill.)
330 and Table Mountain pine (*Pinus pungens* Lamb.) are the dominant pine species in the overstory.
331 Eastern white pine (*Pinus strobus* L.) is the dominant pine species in the overstory in moist
332 coves. A mixture of oaks, including black (*Quercus velutina* Lam.), chestnut (*Quercus montana*
333 L.), northern red (*Quercus rubra* L.), scarlet (*Quercus coccinea* Muenchh.), and white (*Quercus*
334 *alba* L.), are dominant in the overstory throughout. Depending on site characteristics, blackgum
335 (*Nyssa sylvatica* Marsh.), mockernut hickory (*Carya tomentosa* [Poir] Nutt.), red maple (*Acer*
336 *rubrum* L.), sourwood (*Oxydendrum arboreum* [L.] DC), and yellow-poplar (*Liriodendron*
337 *tulipifera* L.) are also present (Waldrop et al. 2010).

338 A mixture of ericaceous shrubs occurs at GRGL, as well, accounting for 35% cover or
339 greater within many plots at the time of study implementation in 2001 (Waldrop et al. 2007).
340 These shrubs include mountain laurel (*Kalmia latifolia*) and blueberry (*Vaccinium* spp. L.)
341 on drier sites, rhododendron (*Rhododendron maximum* L. and *Rhododendron minus* Michx.) in
342 riparian areas and moist coves, and flame azalea (*Rhododendron calendulaceum* [Michx.] Torr.)
343 in submesic to subxeric sites. While native to the region, these species may greatly increase
344 vertical fuel structure, aiding in the formation of ladder fuels that may increase fire intensity and
345 crown mortality in the event of a wildfire under dry conditions (Coates and Ford 2022, Reilly et
346 al. 2022, Waldrop and Brose 1999). They also impede tree regeneration due to the increased
347 shade presented by their dense canopies (Elliott and Miniati 2021, Bolstad et al. 2018, Nilsen et
348 al. 1999).

349 3.2.2 Treatments

350 The southern Appalachian Fire and Fire Surrogate Study was implemented in a
351 randomized complete block design, consisting of four treatments and three replications.

352 Treatments included: mechanical (chainsaw) felling of shrubs and midstory stems less than 10
353 cm diameter at breast height (DBH) (M), prescribed fire (B), a combination of mechanical
354 felling and prescribed fire (MB), and an untreated control (C) (Taylor et al. 2025, Stephens et al.
355 2009). Mechanical felling within the M and MB treatments was conducted at the inception of the
356 study at GRGL, between December 2001 and February 2002 (Dukes et al. 2020). An additional
357 felling was conducted in the M replications in 2012. No additional felling was conducted in the
358 MB replications in 2012 because it was determined they were not necessary (Waldrop et al.
359 2016). Prescribed burns were implemented five times (2003, 2006, 2012, 2015, and 2024) prior
360 to 2025 in both B and MB.

361 3.2.3 Fuels inventory

362 Ten 20 m x 50 m modified Whittaker plots were installed within each of the twelve
363 treatment-replications to conduct vegetation inventories (Waldrop et al. 2016). A modified
364 version (Coates et al. 2019, Coates and Ford 2022) of Brown's Planar Intercept Method (Brown
365 1974) was utilized in each corner of the ten 10 m x 10 m subplots to tally and measure fuels on
366 the ground (n=480). Due to limitations of time and funding in 2025, 25% of the original fuel
367 sampling locations were revisited. Specifics for the layout of the planar intercepts and the
368 formulas used to convert the tallies of down-and-dead woody fuels to volume and mass can be
369 found in Coates et al. (2019) and Coates and Ford (2022). Masses of 1-hr (0-0.64 cm), 10-hr
370 (0.65-2.54 cm), 100-hr (2.55-7.61 cm), and 1000-hr (≥ 7.62 cm) fuels were obtained from the
371 tallies. Depths of litter (Oi Horizon) and duff (Oe+Oa Horizons) were measured using rulers
372 (mm) along the planar intercepts (Coates et al. 2019; Coates and Ford 2022) and were converted
373 to masses using the following depth to mass conversions (Ottmar and Andreu 2007):

374 Litter mass: 1.22 Mg per ha per cm

375 Duff mass: 4.27 Mg per ha per cm

376 For our analyses, we followed the methods of Bernal et al. (2025) and evaluated: 1. The sum of
377 the litter and duff masses as *Ground Fuels* (GF); 2. The sum of the 1-, 10-, and 100-hr fuel
378 masses as *Fine Woody Debris* (FWD); 3. 1000-hr fuel mass as *Coarse Woody Debris* (CWD), 4.
379 The sum of GF, FWD, and CWD as *Total fuel mass* (TFM). Plots that displayed no evidence of
380 wildfire impacts were excluded from our statistical analyses (Figure 3.3).

381 3.2.4 Statistical analyses

382 Means of GF, FWD, CWD, and TFM were determined for 2001 (pre-treatment), 2014,
383 and 2025. The significance of mean fuel load changes for each of the fuel loading categories
384 from 2014-2025 was assessed using a Kruskal-Wallis Test (one-way analysis of variance
385 [ANOVA] for non-parametric data) (Conover, 1984). When differences between treatments were
386 significant ($\alpha=0.05$), a Steel-Dwass All Pairs test was utilized to distinguish those differences.
387 All statistical analyses were conducted using JMP Student Edition 19 (SAS; Cary, NC, USA).

388 3.3 Results

389 Means of the fuel loads in 2001, 2014, and 2025 are listed in Table 3.3 and are visually
390 displayed in Figure 3.4. In 2001, 52-58% of TFM was accounted for by GF in each of the
391 treatments. By 2014, only 18% and 37% of TFM was accounted for by GF in the MB and B
392 treatments, respectively. Masses of GF accounted for 52% of TFM for both the C and M
393 treatments in 2014. In 2025, only 4-9% of TFM was accounted for by GF in all treatments. In all
394 years, FWD accounted for 8-27% of TFM. Loads of CWD were 25-38% of TFM in 2001. By
395 2014, 43% and 67% of TFM was accounted for by CWD in the B and MB treatments,
396 respectively, while CWD remained nearly constant in the C treatment at 29%. In the M

397 treatment, 23% of TFM in 2014 was accounted for by CWD, a decrease of 15% from 2001. In
398 2025, CWD loads accounted for 64-78% of all TFM loads in all treatments.

399 From 2014-2025, the mean reduction in GF loads differed between treatments ($p < 0.01$)
400 (Table 3.2) and was greatest in the C treatment. This reduction was not significantly different
401 from the M treatment, but it did differ from the B and MB treatments. No significant differences
402 between the means of the treatments existed for FWD ($p = 0.19$, CWD ($p = 0.06$), or TFM ($p = 0.33$)
403 (Table 3.2). However, while not statistically significant: 1. CWD increased in both the C and M
404 treatments, but decreased in the B and MB treatments; 2. TFM decreased in all treatments except
405 the M treatment from 2014-2025; 3. In 2025, TFM was 26.93 Mg ha^{-1} in the B treatment,
406 representing only 53-61% of TFM in the other treatments.

407 3.4 Discussion

408 3.4.1 Impacts of treatments prior to 2024

409 As previously stated by Bernal et al. (2025) and as shown in Figure 3.4 (Table 3.1), the
410 implementation of the B and MB treatments reduced GF on the GRGL from 2001-2014. The
411 reduction in GF, which serves as the primary fuel source for mixed oak-pine forests in the
412 eastern United States (Waldrop and Goodrick 2016), was the result of repeated prescribed fires.
413 The implementation of these fires in these long-unburned stands caused delayed mortality of
414 overstory stems (Taylor et al. 2025, Waldrop et al. 2016), leading to subsequent increases in
415 CWD ($>200\%$ in B, $>400\%$ in MB). Changes in overstory light conditions promoted desired tree
416 regeneration, with the most positive outcomes aligning with the MB treatment, followed by the
417 B treatment (Taylor et al. 2025). These positive outcomes steered management objectives toward
418 continued burning at the site prior to 2024. Potential innovations of the treatments were
419 discussed prior to October 2024, such as the use growing season prescribed fires (Vaughan et al.
420 2022) and experimentation with herbicides. These innovations would be targeted to reduce

421 ericaceous shrub cover and constrain regeneration of red maple and other less desired tree
422 species.

423 3.4.2 Impacts of Hurricane Helene and the Black Cove Wildfire Complex

424 In October 2024, plans for potential next steps at GRGL were paused due to the impacts
425 of Hurricane Helene. Several months later in March 2025, the Black Cove Wildfire Complex
426 further confounded future planning. Impacts from both Hurricane Helene and the Black Cove
427 Wildfire Complex varied widely between plots within treatments, and between treatments within
428 replications (Figure 3.5). Overstory blowdowns throughout GRGL were patchy in distribution,
429 with some areas seeing nearly 100% of overstory trees toppled by storm winds and others seeing
430 0%. In some cases, the fuel transects were oriented within patches of forest apparently
431 unaffected by the storm and completely excluded from the wildfire. In others, transects were
432 placed within dense, “jack-strawed” timber. In some of the more severely burned sites, no O
433 Horizon was present, whereas in unburned areas of the M, B, and MB treatments, litter depth
434 sometimes matched or exceeded pre-treatment levels. The spread of the Black Cove Wildfire
435 Complex, largely unchecked within the bounds of natural and constructed control features,
436 immensely complicated this controlled experiment as it crossed FFSS treatment boundaries.

437 Changes in mean fuel load from 2014-2025 differed by treatment for GF (Table 3.2). This
438 reduction was greatest for C, but did not differ significantly from the M treatment. This result is
439 logical because fire had not been present in these sites previously. Fire ignition and spread
440 through those treatment areas was aided by fuel complexes that were likely 52% GF, based upon
441 the 2014 values (Table 3.1). Due to the high degree of variability of impacts from Hurricane
442 Helene and the high variability in fuel loads within each sampling location, the CWD change
443 was not significant between the treatments. However, CWD decreased in the B and MB

444 treatments, but increased in the C and M treatments. In the B and MB treatments, some of the
445 CWD that had been added as a result of overstory mortality induced by the treatments prior to
446 2024 was likely consumed in the wildfire. In the C and M treatments, new CWD was likely
447 added as a direct result of Hurricane Helene, the Black Cove Wildfire Complex, or both events.
448 Although both treatments experienced reductions in CWD in 2025, the remaining CWD in the
449 MB treatment was double that of the B treatment.

450 3.4.3 Fuels and wildfire management

451 While we have no documented photographs or measurements of fire behavior within
452 each plot, we hypothesize that fire behavior likely differed in the B and MB treatments from the
453 C and M treatments due to the differences in both fuel loading and fuel structure. This brings to
454 light a critical consideration for fuel treatments: TFM may stay constant despite changes in fuel
455 structure. When fuel structure is altered, both vertical and horizontal fuel continuity may be
456 impacted. Alterations to horizontal continuity may impact surface fire spread, and alterations to
457 vertical continuity may impact ladder fuels and the potential for crown fires (Keane 2015). When
458 the fuel reduction treatments were coupled with another natural disturbance, such as Hurricane
459 Helene, overstory stems that may have maintained resilience to repeated burning may have
460 succumbed to inundation with heavy rains and strong winds. Additional work is being conducted
461 to further investigate those dynamics specifically at GRGL (Weise 2026, in preparation).

462 From a practical standpoint of wildfire management following the fuel reduction
463 treatments, we do know that the Black Cove Wildfire Complex did not fully spread throughout
464 all of the study area (Figure 3.5). For example, in replication 1, the wildfire only spread through
465 a portion of the B treatment. In replicate 3, the majority of the M treatment was unaffected. All
466 of replication 2 was part of the wildfire complex, regardless of treatment designation.

467 Unfortunately, we are not certain of the points of ignition nor the tactical decisions that were
468 made to constrain the boundaries of the wildfire. Wildland firefighters often make tactical
469 suppression decisions to insure the protection of lives and property, therefore we cannot directly
470 state that some of the treatments promoted more or less spread of the wildfire. However, we have
471 evidence that the fuel reduction treatments did not provide “wildfire-proof conditions” at GRGL.
472 The notion that fuel mitigation treatments completed remove any risk of wildfires is a faulty
473 assumption and further affirms that changes in fire behavior may be expected following
474 alterations to fuel loading and structure rather than complete wildfire resistance. Fuels mitigation
475 in the southern Appalachian Mountains may target the reduction of hazardous wildfire conditions
476 to reduce threats to lives and property if and when wildfires occur, but the potential for wildfires
477 in systems that are managed with prescribed fires will still be present under certain conditions
478 because fuels must be present to ignite and sustain a prescribed fire regime. Areas managed with
479 prescribed fire benefit from the fuel breaks and fire perimeters that are often created and
480 maintained to sustain frequent prescribed burning, and those assets may provide benefits when
481 and if wildfires occur in areas that have been treated with prescribed fires.

482 3.5 Conclusion

483 Next steps for GRGL began in May 2025. A prescribed fire was administered in that
484 month to include all locations within replication 1 that had not been impacted by the wildfire. We
485 anticipate the installation of additional treatments at the site to target objectives for both wildfire
486 risk reduction and the long-term management of a mixed oak-pine forest. The additional
487 treatments may include herbicides and the use of prescribed fire in the growing season to
488 potentially alter fuel composition and subsequent fire behavior. In the southern Appalachian
489 Mountains, prescribed fire is being used more frequently to target a variety of objectives,

490 including but not limited to fuel hazard reduction, vegetation restoration and maintenance, and
491 wildlife habitat restoration and maintenance (Waldrop and Goodrick 2012). Many locations
492 within the region have now been burned numerous times. Findings from GRGL of use to
493 managers include: 1. Long-term use of B and MB treatments does not ensure that a forest will
494 resist impacts from disturbances such as wildfires, windstorms, and hurricanes; and 2.
495 Achievement of long-term land management objectives may require multiple entries of multiple
496 treatments over time.

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Roberto Di Giovine | U.S. Army National Guard via AP



AP Photo / Allison Rouse



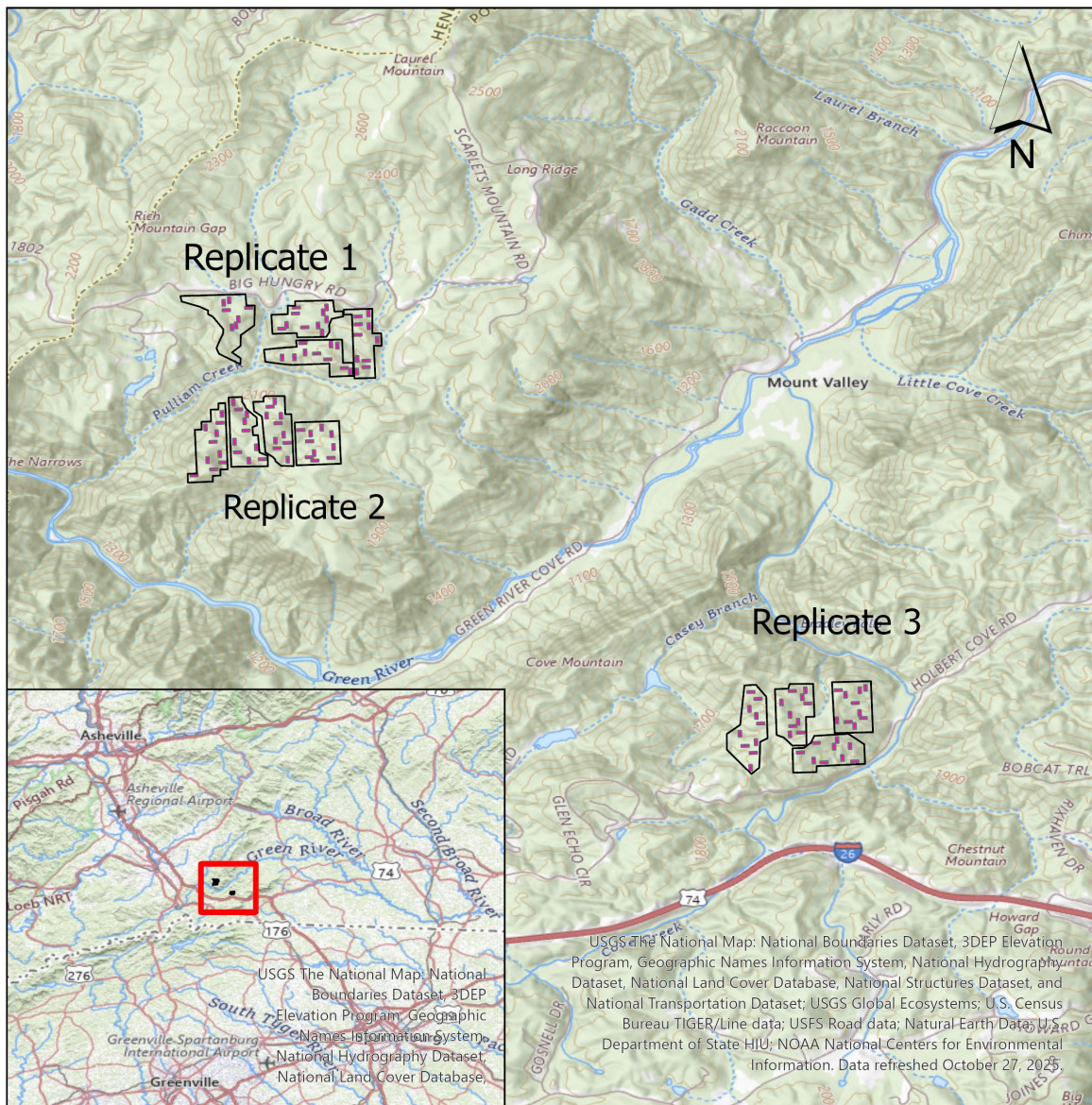
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Figure 3.1 The Black Cove Wildfire Complex (March 2025) burned approximately 3,076 ha in Henderson and Polk Counties, North Carolina.

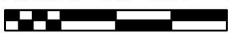
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Green River Game Land, North Carolina Research Area



Legend

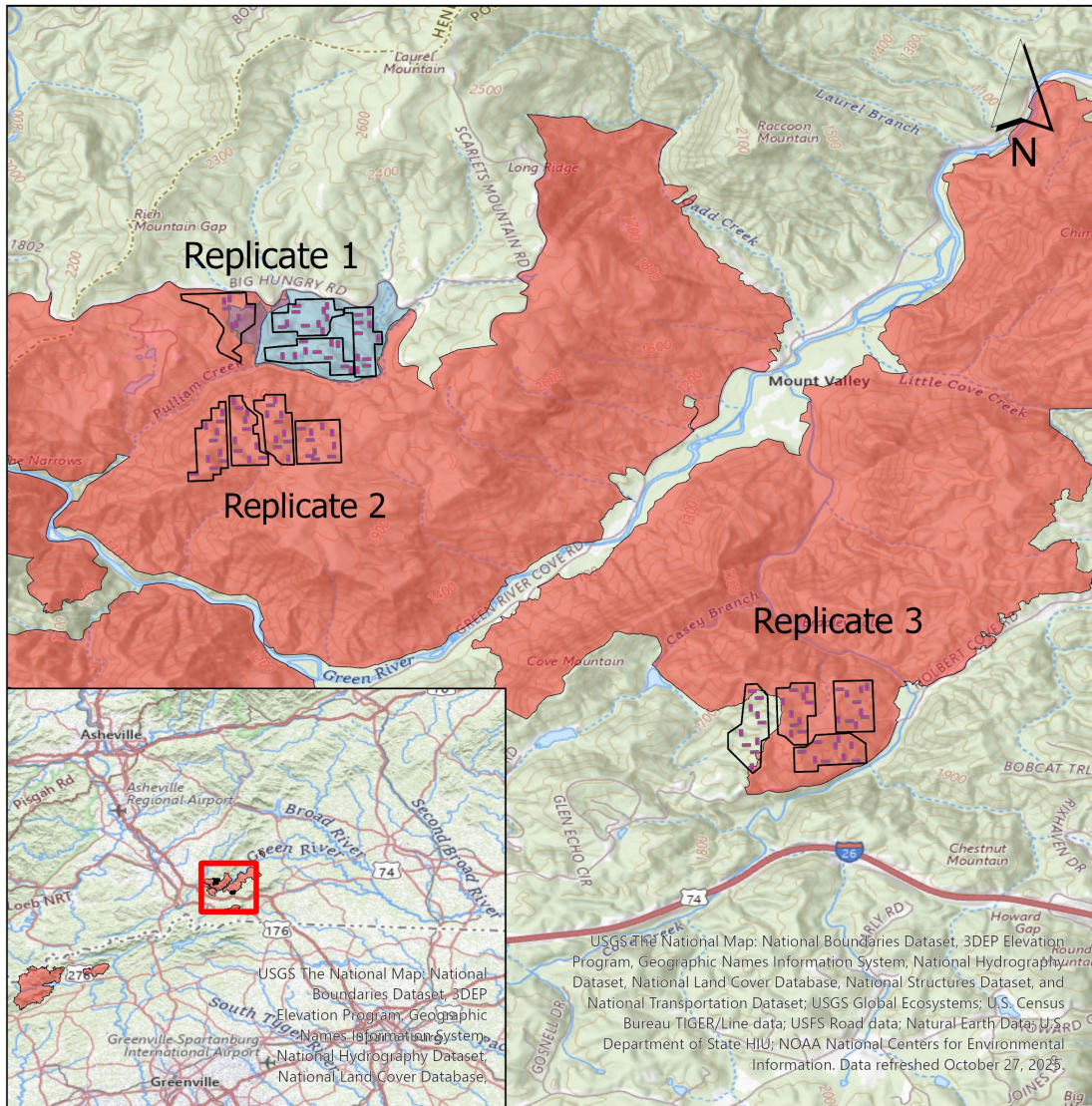
- Treatment Boundary
- FFSS Plots

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Figure 3.2 Location of the Southern Appalachian Fire and Fire Surrogate Study on the Green River Game Land near Hendersonville, North Carolina, USA

Black Cove Wildfire Complex Perimeter



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Figure 3.3 Extent of impact of the Black Cove Wildfire Complex within the Southern Appalachian Fire and Fire Surrogate Study, Green River Game Land, North Carolina, USA. Note that three treatment units in Replication 1 and one treatment unit within Replication 3 were not completed impacted by the wildfire.

529
530 **Figure 3.4** Means (\pm standard error of the mean) for the masses of ground fuels (GF), fine woody
531 fuels (FWD), coarse woody debris (CWD), and total fuels (TF) in 2001, 2014, and 2025 for each
532 of the treatments (C = control, B = burn-only, M = mechanical-only, and MB =
533 mechanical+burn) of the Southern Appalachian Fire and Fire Surrogate Study, North Carolina,
534 USA. Plots not observably impacted by wildfire were excluded from this analysis.
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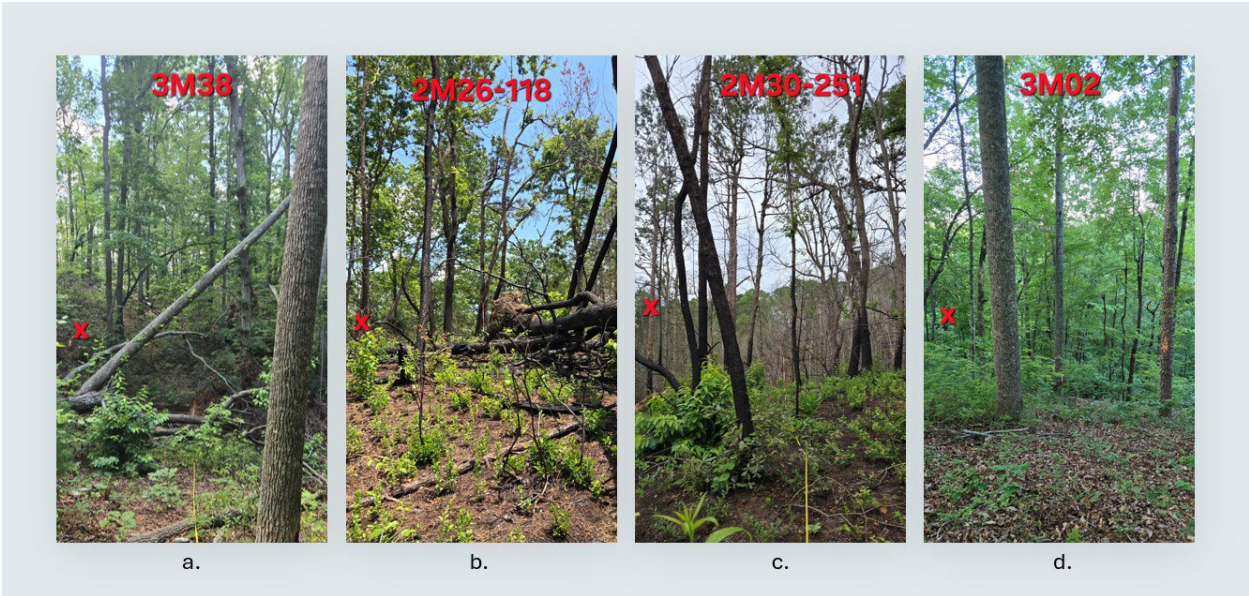


Figure 3.5 Variable impacts of Hurricane Helene and the Black Cove Wildfire Complex at the Green River Game Land, North Carolina, USA as observed in June-July 2025. This panel depicts mechanical-only treatments under four observed conditions: a.) impacted by the storm; b.) impacted by the storm and the wildfire; c.) impacted only by the wildfire; d.) not significantly impacted by the storm and excluded from the wildfire.

584 **Table 3.1a – 3.1c** Means (\pm standard error of the mean) for the fuel loading variables in 2001,
 585 2014, and 2025 for each of the treatments of the Southern Appalachian Fire and Fire Surrogate
 586 Study, North Carolina, USA. The percentage of total fuel mass represented by ground fuels, fine
 587 woody fuels, and coarse woody debris is listed below the mean (\pm standard error of the mean).
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589 **Table 3.1a** 2001

Fuel Mass (Mg ha⁻¹)	Control (n=19)	Burn-only (n=22)	Mechanical- only (n=12)	Combination (n=16)
Ground Fuels	21.04 (1.74) 58%	23.78 (1.71) 58%	18.66 (1.75) 54%	20.00 (1.92) 52%
Fine Woody Fuels	5.77 (0.75) 16%	6.97 (0.94) 17%	2.89 (0.53) 8%	6.35 (0.91) 17%
Coarse Woody Debris	9.75 (2.27) 27%	10.19 (3.27) 25%	12.94 (4.30) 38%	11.96 (4.18) 31%
<i>Total</i>	36.56 (2.71)	40.93 (3.25)	34.49 (4.63)	38.31 (5.46)

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Table 3.1b 2014

Fuel Mass (Mg ha⁻¹)	Control (n=19)	Burn-only (n=22)	Mechanical- only (n=12)	Combination (n=16)
Ground Fuels	26.00 (1.40) 52%	19.31 (1.11) 37%	23.28 (1.96) 52%	13.86 (1.13) 18%
Fine Woody Fuels	9.36 (1.80) 19%	10.39 (0.81) 20%	11.18 (1.10) 25%	10.87 (1.26) 14%
Coarse Woody Debris	14.53 (3.35) 29%	22.57 (5.36) 43%	10.55 (2.21) 23%	50.40 (14.90) 67%
<i>Total</i>	49.90 (3.94)	52.27 (5.69)	45.01 (3.54)	75.14 (15.24)

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Table 3.1c 2025

Fuel Mass (Mg ha⁻¹)	Control (n=19)	Burn-only (n=22)	Mechanical- only (n=12)	Combination (n=16)
Ground Fuels	2.24 (0.19) 5%	2.53 (0.52) 9%	3.70 (0.69) 7%	1.92 (0.38) 4%
Fine Woody Fuels	7.97 (0.85) 18%	7.15 (0.71) 27%	7.51 (0.92) 15%	8.65 (1.41) 19%
Coarse Woody Debris	33.78 (10.75) 77%	17.25 (4.36) 64%	39.39 (15.40) 78%	34.76 (11.36) 77%
<i>Total</i>	43.98 (10.94)	26.93 (4.35)	50.60 (16.11)	45.32 (12.30)

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Table 3.2 Mean fuel mass change (\pm standard error of the mean) from 2014-2025 for each of the treatments of the Southern Appalachian Fire and Fire Surrogate Study, North Carolina, USA. Negative values indicate fuels were added over the time period.

Fuel Mass Change (Mg ha⁻¹)	Control (n=19)	Burn-only (n=22)	Mechanical-only (n=12)	Combination (n=16)	p-value
Ground Fuels	23.76 (1.37) A	16.77 (1.01) B	19.58 (1.93) AB	11.94 (0.97) C	<0.01
Fine Woody Fuels	1.40 (1.90)	3.24 (1.02)	3.67 (1.25)	2.23 (2.01)	0.19
Coarse Woody Debris	-19.25 (11.29)	5.32 (4.66)	-28.84 (14.95)	15.64 (11.55)	0.06
<i>Total</i>	5.92 (11.93)	25.34 (5.33)	-5.59 (16.47)	29.81 (12.61)	0.33

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631 **Chapter 4. Conclusion**

632 4.1 Fuels summary

633 As the climate changes, and the frequency and severity of storms like Hurricane Helene
634 are projected to increase, effective management for fire-adapted and resilient forests becomes
635 even more vital. Managers must seek not only to mitigate the results of broad-scale fire
636 exclusion in restoring our forests' natural resilience, but to increase that resilience to a level
637 sufficient to withstand higher temperatures and more frequent and intense events. Therefore, this
638 thesis represents timely work addressing a topic of interest to many within the Appalachian
639 Mountain region.

640 Results from the study of post-Helene, post-Black Cove Wildfire fuels at GRGL suggest
641 that long-term alteration of both fuel loading and fuel structure is challenging. Areas receiving
642 the fuel reduction treatments were not invincible to wildfire ignition and spread. While we have
643 no recorded evidence to illustrate differences in fire behavior during the Black Cove Wildfire
644 Complex in the B and MB treatments versus the C and M treatments, it appears likely that fire
645 behavior differed between these treatments based upon the differences in the fuel components
646 that would have been present prior to the wildfire and their distribution. Similarly, we cannot be
647 100% certain of the tactical fire suppression decisions that may have encouraged or stifled fire
648 spread and residence time in each treatment-replicate.

649 4.2 Additional preliminary analyses

650 4.2.1 Hurricane Helene impacts

651 Additional work is needed to fully characterize the impacts of the fuel reduction
652 treatments on hurricane resistance at GRGL. Data from Christie and Norman's (2024) HiForm
653 Hurricane Helene Rapid Assessment datasets were used to perform a cursory analysis of storm

654 damage severity at GRGL. The HiForm assessment used 10 m Sentinel-2 imagery to determine
655 NDVI change, and this was combined with documentation from field observations to assess
656 storm damage to forests and other lands along the length of Helene’s impact area, from its
657 landfall in the Florida Panhandle to its eventual dissipation in southern West Virginia. Impact
658 severity was classified into four types:

- 659 • Type 1: Extreme NDVI change (≤ -20) and in patches greater than 0.1 ha; typically
660 associated with complete stand blowdown, or partially blown down stands with a major
661 increase in CWD.
- 662 • Type 2: Moderate NDVI change (≤ -13); may include small gaps of extreme blowdown
663 ≤ -20 .
- 664 • Type 3: Low NDVI change (-6 to -12); normally an indication of isolated gaps and/or
665 minor crown damage, may also represent leaf stripping from wind.
- 666 • Type 4: No or minor impacts from the storm.

667 We overlaid HiForm’s raster dataset with the GRGL raster and plot map in ArcGIS Pro, and a
668 spatial join was used to associate storm damage severity with individual plot locations. Data
669 from this join function was then extracted for analysis.

670 We found that most Hurricane Helene impacts to plots at GRGL (54%) were classified as
671 Type 4, with little or no apparent damage (Table 4.1). Impacts to 31% of plots were classified as
672 Type 3, displaying isolated gaps and/or minor crown damage at the time of assessment. The most
673 severe impacts within GRGL treatments were classified as Type 2, and impacted 15% of the
674 study area. Thirty-two percent of MB plots had Type 2 damage, and 56% percent of plots with
675 Type 2 damage were MB. No Type 1 impacts were located within the study area at GRGL. For
676 the purposes of this study, the term “storm-impacted” shall refer to plots assessed at impact

677 severity Type 2. Therefore, 15% of the GRGL site is considered “storm-impacted” as pertains to
678 fuel loading.

679 At GRGL, damage severity was greatest within treatments with the lowest overstory
680 density and least in treatments with the greatest overstory density ($p < 0.0001$). The MB
681 treatment was broadly occupied by early successional forest vegetation in late 2024 and saw the
682 highest incidence of high severity impacts. Thirty-two percent of MB plots had Type 2 damage,
683 and 56% percent of plots with Type 2 damage were MB. The C treatment remains mostly
684 mature, closed canopy forest, which experienced little to no severe storm damage.

685 4.2.2 Black Cove Wildfire Complex impacts

686 Impacts of the Black Cove Wildfire Complex at GRGL were preliminarily assessed by
687 overlaying the North Carolina Forest Service (NCFS) fire progression map with our treatment
688 and plot map in ArcGIS Pro. A spatial join was then used to classify plots as either burned or not
689 burned based on their inclusion or exclusion from the wildfire perimeter, and this classification
690 was checked against field observations of conditions at each plot. The progression map provided
691 the dates during which specific areas were impacted by the wildfire, allowing extraction of
692 substantial information regarding weather and fire danger conditions at the time of impact for
693 each plot (Table 4.2).

694 Following the wildfire, in May 2025, efforts were made to complete the wildfire’s work
695 in some of the areas excluded by suppression operations. The majority of Replication 1, which
696 had been excluded from the Black Cove Wildfire Complex, was treated with low intensity
697 prescribed fire on May 6, 2025, under less volatile weather conditions than occurred during the
698 Black Cove Wildfire Complex. In addition to burn status, each plot was classified by burn type.
699 Plots confirmed to be impacted by wildfire were classified with burn type WF in our dataset,

700 while those within the prescribed burn perimeter, and confirmed to have been impacted by that
701 burn, were classified with burn type RX.

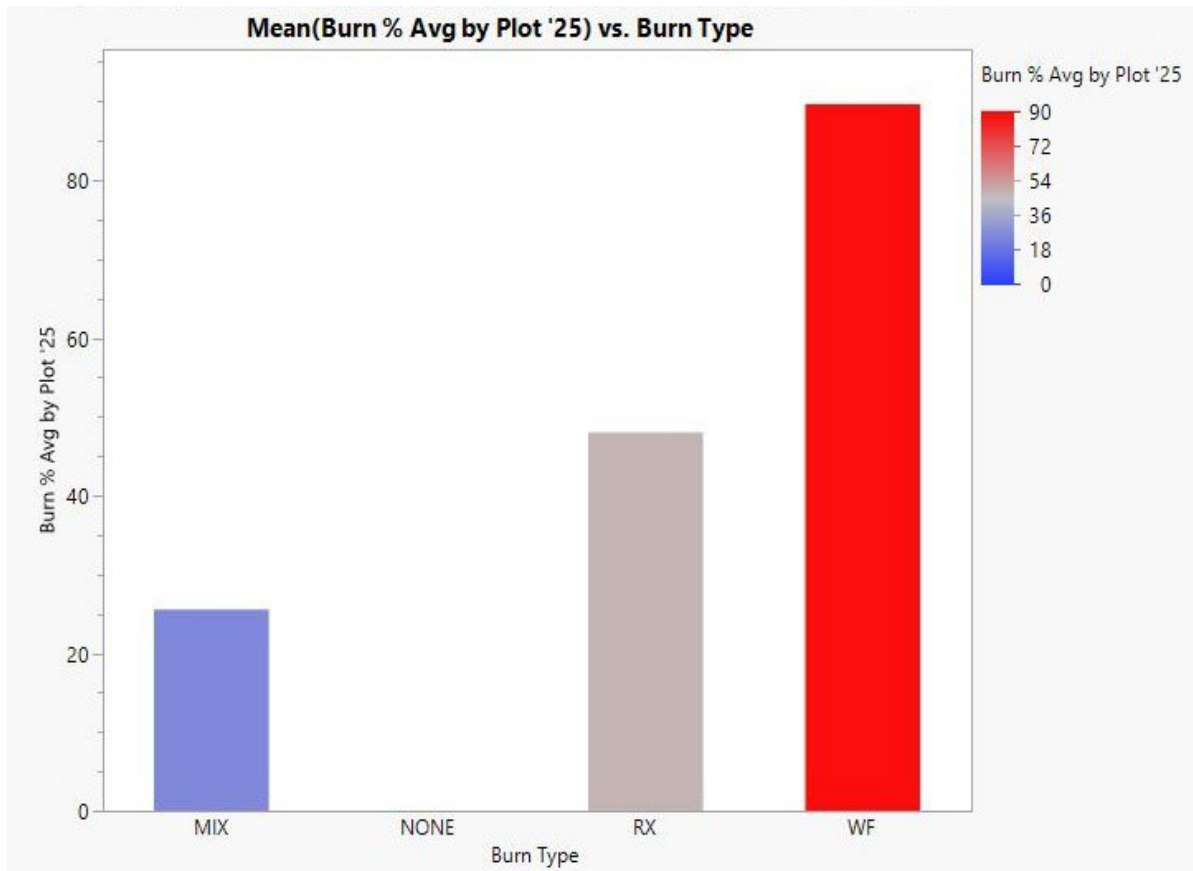
702 The SDA Fire Environment Mapping System (FEMS) provides fire weather and fire
703 danger data from over 2,000 Remote Autonomous Weather Stations (RAWS) across the United
704 States (Wildfire.gov). For fire weather and fire danger analysis, we selected the Guion Farm
705 RAWS, located approximately 24.1 km from the site, in Hendersonville, NC. Fire weather and
706 fire danger data were extracted for each date during which our plots were impacted by fire, and
707 that data was associated with each plot based on its date of burn. To facilitate percentile analysis
708 of these metrics, we extracted hourly weather and fire danger data from the station for every day
709 of every year since 2005. This data was used to determine 80th percentile weather conditions in
710 keeping with those considered for FFSS objectives (Stevens 2009). The same 80th percentile
711 threshold was considered for common National Wildfire Coordinating Group (NWCG) fire
712 danger metrics at each site.

713 Among the plots measured, 89% were confirmed to have experienced either wildfire or
714 prescribed fire in 2025 (Figure 4.2, Table 4.3). Sixty-four percent of fire-impacted plots were
715 confirmed to have burned by wildfire, while 25% were burned by the May 2025 prescribed fire.
716 Of the remaining plots, 8% experienced no fire at all, and 3% of all plots were observed to have
717 been partially burned. Fuel transects in partially burned plots generally did not include a
718 significant portion of the burned area within these plots.

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733 **Figure 4.2** The percentage of plot area impacted by burn type the Green River Game Lands,
 734 Hendersonville, North Carolina, USA. 'WF' represents plots burned by wildfire. 'RX' represents
 735 plots burned only by prescribed fire. 'MIX' represents plots partially burned by wildfire.

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751 **Table 4.1** The percentage of plots located at the Green River Game Land (Hendersonville, NC,
752 USA) affected by the storm severity types produced by Hurricane Helene (HiForm Hurricane
753 Helene Rapid Assessment - Christie & Norman, 2024)
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Storm Severity Type	Percentage of Plots Impacted (%) n=111
1 (Highest)	0
2	15
3	31
4 (Lowest)	54

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792 **Table 4.2** Mean fire weather and fire danger values for the Green River Game Land during the
 793 Black Cove Wildfire Complex (March 2025) and the May 2025 prescribed fire. Values meeting
 794 the Fire and Fire Surrogate Study 80th percentile thresholds are shown in red.
 795

Fuel/Weather Parameter	March 22	March 24	March 28	March 29	May 6
Keetch-Byram Drought Index (KBDI)	55.00	70.33	102.00	111.33	22.33
Energy Release Component (ERC)	31.00	26.02	35.20	30.80	13.09
Burning Index (BI)	19.03	12.17	19.30	16.07	8.03
Spread Component (SC)	1.82	1.03	1.64	1.24	0.88
1-hr Fuel Moisture (%)	7.98	17.09	9.34	14.88	19.29
10-hr Fuel Moisture (%)	9.65	16.17	9.80	13.45	17.43
100-hr Fuel Moisture (%)	13.87	12.97	12.34	12.38	18.77
Live Woody Fuel Moisture (%)	60.00	69.67	87.67	93.33	174.00
Live Herb Fuel Moisture (%)	30.67	46.33	73.67	83.00	209.33
Temperature (°C)	10.83	13.73	14.72	15.63	12.80
Relative Humidity (%)	27.04	56.17	35.88	60.50	69.54
Wind Gust Speed (km h ⁻¹)	19.52	19.44	14.89	18.51	11.88

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805 **Table 4.3.** The percentage of plots impacted by wildfire or prescribed fire in 2025 on the Green
806 River Game Land, Hendersonville, NC, USA.
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Burn Status and Type	Percentage of Plots Affected (%) n=111
Burned	89
Not Burned	8
Partially Burned	3
Burned by Prescribed Fire	25
Burned by Wildfire	64

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